# Review and Selection of Fish Metrics <br> for the Chicago Area Waterway System Habitat Evaluation and Improvement Study 

Prepared for

The Metropolitan Water Reclamation District Of Greater Chicago

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## LIST OF ATTACHMENTS

Attachment A: CAWS Fish Data Stations and Sampling Dates Used in this Analysis
Attachment B: List of Fish Species Identified in the CAWS (2001-2007) and Their Tolerance Assignments

Attachment C: Matrix of Pearson's Correlation Values for Fish Metrics

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## 1. INTRODUCTION

This document summarizes the process used to select key fish metrics for the Chicago Area Waterway System (CAWS) Habitat Evaluation and Improvement Study (the Study). Selection of key fish metrics is important to the Study for two reasons:

- Sensitivity to Water Quality - Comparison of historical fish data and water quality data is important in understanding the relationship between water quality and fish communities in the CAWS. Identification of CAWS appropriate fish metrics is necessary for such data comparisons.
- Habitat Index Development - The proposed method for development of a CAWSspecific habitat index relies on the comparison of fish data to habitat variables to help define the relationship between fish and the physical habitat in the CAWS.

It was not the objective of the Study to develop a CAWS-specific index of biotic integrity (IBI), but the methods used to identify key fish metrics for the CAWS are the same as those used in current biological practice to define metrics for fish IBIs. Development of a fish IBI for the CAWS might be useful in the future, but development an IBI would require specification of a regionally appropriate, non-consumption, target condition to which the upper end of the index would be referenced (Karr 1991). This can be done in one of three ways, but is currently beyond the scope of this analysis for the CAWS as described below:

- External reference reach - An external reference reach that represents a target fisheries condition that is attainable in the CAWS could be used to establish the upper limit of the IBI. This approach is impractical for the CAWS because the CAWS consists entirely of constructed or heavily modified channels and no similar channels with high quality or reference fisheries have been identified.
- Internal reference reach - A reach within the system that represents a target fisheries condition that should be targeted for the entire CAWS could be used to establish the upper limit of the IBI. This is not currently possible because no such internal reference has been identified.
- Target use - A target fisheries use (e.g., warm water sport fishing), function (e.g., harvest prohibition) or specific target species (e.g., trophy largemouth bass) may be identified which would allow determination of target fisheries conditions to describe the upper end of the index. To date, target uses or species have not been identified.

Although it is currently impractical to establish a fish IBI for the CAWS, it is possible to determine key fish metrics for use in comparing to habitat data. This document presents the recommended list of fish metrics for the CAWS and summarizes the methodology used to arrive at that list.

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## 2. DATA DESCRIPTION

This section provides an overview of the fish data used in this study.

### 2.1 FISH SAMPLING LOCATIONS

The Metropolitan Water Reclamation District of Greater Chicago (District) has been collecting fish data annually since 1974 (with the exception of 1981 and 1982) within the CAWS. However, to focus this Study on current conditions, LimnoTech limited the fish data analysis to the data collected between 2001 through 2007 and to the area considered as the managed portion of the CAWS. The managed portion is defined by the nonwadeable waters bounded by the Wilmette Pumping Station, the Chicago River Lock and Controlling Works, the O’Brien Lock and Controlling Works and the Lockport Lock and Powerhouse. The tributaries to the CAWS are not included in this study, as their physical conditions and regulatory controls differ from the mainstems of the CAWS. The South Fork of the South Branch, also known as Bubbly Creek, is also included in this study.

During the 2001-2007 period, the District collected fish data at 34 stations within the CAWS (Figure 2.1) on a routine basis. Twenty-six of these 34 stations are part of the District's Ambient Water Quality Monitoring (AWQM) program. Seven of the AWQM stations are annually monitored (once per year), while the remainder are sampled on a four year rotation. The total number of sample events across all stations and years includes 113 sample events. The CAWS fish monitoring stations and sampling dates used in the sample description, screening and selection of fish metrics is included as Attachment A.

### 2.2 FISH SAMPLING METHODS

The District samples the fishery within the CAWS using boat electrofishing procedures ${ }^{1}$, following standard and consistent protocols for this collection method. Each station is generally defined by a 400 meter reach and each bank length was sampled for fishes. The average shock time averages 800 seconds. The collected fish are counted, measured (standard and total length), weighed and released, except where difficult to identify in the field. In addition, any abnormalities such as diseases, eroded fins, lesions or tumors (DELTs) are noted. Between 2001 through 2007, all sampled stations have a single sampling event per year except Station 75, Chicago Sanitary and Ship Canal at Cicero Avenue. During the first sampling event on 7/31/2001 the field crew experienced equipment failure, which resulted in a partial fish collection sample. Later in the season, on $9 / 4 / 2001$ the crew returned to the station to conduct an additional sampling. Only the $9 / 4 / 2001$ data were included in this study. Finally, supplemental sampling was conducted in 2007 using Fyke nets at three stations, and those data are also summarized.

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### 2.3 SUMMARY DESCRIPTION OF FISH DATA

Fifty-two (52) species, including five hybrids, of fish were identified at the 34 CAWS monitoring stations between 2001 and 2007 (sample period). Attachment B provides the complete list of these fish species. For the sample period, the number of non-hybrid species collected across the CAWS stations ranged from 27 at AWQM Station 76 (Little Calumet River at Halsted Street) to only five at Stephen Street (Chicago Sanitary Shipping Canal; CSSC; Figure 2-2). The repeated, annual sampling effort did not necessarily relate to the greatest number of taxa among the sample period for an individual station. For example, the second most numerous taxa ( $n=23$ ), were from the Little Calumet River at Indiana Avenue, resulting from only two sample events for the sample period. Figure 2-1 depicts the distribution of the number of non-hybrid collected taxa across the managed portion of the CAWS. Table 2-1 describes the taxa richness and total number of individuals by station, for the sample period.


Figure 2-1. Fish Sampling Stations in the CAWS.


Figure 2-2. Taxa Collected among CAWS Stations for the 2001-2007 sample period. Blue bars indicate stations included in the quadrennial sampling schedule while the orange bars indicate those sampled annually.

Table 2-1 summarizes the station sample collections by station and year. Station sampling within the CAWS has ranged from as few as 12 stations (2001) to as many as 20 stations (2005) with an average of 16 stations sampled per year. Station samples vary in their taxa and total number of individuals both within stations among years, and among stations. The least number of species collected in any event occurred in 2001 at Lockport with only 2 taxa represented by 77 individuals. The greatest number of species for a single event included 22 taxa represented by 405 individuals collected on the Little Calumet River at Halsted Street in 2006.

Table 2-1. Taxa Richness and Total Number of Individuals by Station and Year.

| AWQM Station Number | Station Description | Sample taxa richness (total number of individuals) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 35 | North Shore Channel at Central Street | 12 (132) |  |  |  | 11 (139) |  |  |
| 36 | North Shore Channel at Touhy Avenue | 11 (596) | 12 (147) | 14 (335) | 11 (249) | 9 (276) | 16 (496) | 14 (387) |
| 101 | North Shore Channel at Foster Avenue | 15 (179) |  |  |  | 17 (273) |  |  |
| 102 | North Shore Channel at Oakton Street | 2 (2) |  |  |  | 17 (151) |  |  |
| 37 | North Branch Chicago River at Wilson Avenue | 9 (75) |  |  |  | 11 (122) |  |  |
| 73 | North Branch Chicago River at Diversey Parkway | 7 (58) |  |  |  | 13 (164) |  |  |
| 56 | Little Calumet River at Indiana Avenue |  |  | 17 (452) |  |  |  | 18 (322) |
| 76 | Little Calumet River at Halsted Street | 16 (210) | 17 (163) | 13 (219) | 17 (207) | 19 (913) | 22 (405) | 21 (281) |
| SEPA2 | Little Calumet River at SEPA 2 |  |  |  |  | 16 (529) | 12 (218) |  |
| 43 | Calumet-Sag Channel at Route 83 |  |  | 7 (43) |  |  |  | 9 (261) |
| 58 | Calumet-Sag Channel at Ashland Avenue |  |  | 13 (95) |  |  |  | 12 (131) |
| 59 | Calumet-Sag Channel at Cicero Avenue | 10 (127) | 13 (174) | 12 (56) | 10 (147) | 10 (453) | 15 (214) | 12 (297) |
| SEPA3 | Calumet-Sag Channel at SEPA 3 |  |  | 13 (148) |  | 16 (253) |  | 14 (407) |
| SEPA4 | Calumet-Sag Channel at SEPA 4 |  |  | 11 (93) | 11 (82) | 14 (663) | 9 (79) | 15 (417) |
| SEPA5 | Calumet-Sag Channel at SEPA 5 |  |  | 12 (232) | 7 (41) | 16 (443) | 7 (37) | 17 (216) |
| Supplemental Survey | Calumet-Sag Channel at 104th Street |  |  |  |  |  |  | 10 (92) |
| Supplemental Survey | Calumet-Sag Channel at Kedzie Avenue |  |  |  |  |  |  | 8 (87) |
| Supplemental Survey | Calumet-Sag Channel at Southwest Highway |  |  |  |  |  |  | 13 (127) |
| 46 | North Branch Chicago River at Grand Avenue | 12 (53) | 7 (28) | 8 (67) | $9(88)$ | 5 (77) | 10 (158) | 13 (117) |
| 74 | Chicago River at Lake Shore Drive |  | 8 (22) |  |  |  | 7 (83) |  |
| 100 | Chicago River at Wells Street |  | 11 (136) |  |  |  | 10 (250) |  |
| 39 | South Branch Chicago River at Madison Street |  | 10 (138) |  |  |  | 6 (99) |  |
| 40 | Chicago Sanitary and Ship Canal at Damen Avenue |  | 10 (148) |  |  |  | 12 (164) |  |
| 99 | Bubbly Creek at Archer Avenue |  | 5 (21) |  |  |  | 13 (156) |  |
| 108 | South Branch Chicago River at Loomis Street |  | 10 (76) |  |  |  | 13 (142) |  |
| 99.2 | Bubbly Creek at 35th St. |  |  | 5 (39) | 8 (27) | 5 (26) |  |  |
| 99.1 | Bubbly Creek at I-55 |  |  | 6 (31) | 10 (60) | 5 (31) |  |  |
| 99.3 | Bubbly Creek at RAPS |  |  | 7 (151) | 10 (97) | 5 (62) |  |  |
| 41 | Chicago Sanitary and Ship Canal at Harlem Avenue | 9 (88) | 11 (188) | 10 (225) | 13 (193) | 14 (758) | 15 (388) | 12 (282) |
| 42 | Chicago Sanitary and Ship Canal at Route 83 |  | 5 (32) |  |  |  | 5 (10) |  |
| 48 | Chicago Sanitary and Ship Canal at Stephen Street |  | 4 (24) |  |  |  | 5 (24) |  |
| 75 | Chicago Sanitary and Ship Canal at Cicero Avenue** | 10 (118) | 10 (136) | 9 (138) | 13 (191) | 7 (184) | 11 (205) | 13 (280) |
| 92 | Chicago Sanitary and Ship Canal at Lockport (16th Street) | 2 (77) | 6 (67) | 7 (67) | 4 (22) | 9 (179) | 8 (64) | 6 (64) |
| SEPA5_CSSC | Chicago Sanitary and Ship Canal at SEPA 5 |  |  | 5 (18) | 8 (53) | 6 (306) | 8 (34) | 9 (178) |

Figure 2-3 depicts the sample variation among years at the annual stations. The figure also includes the annual variation of species assigned to pollution tolerance categories of tolerant (to pollution), intolerant and moderately tolerant. A discussion of the categorical assignments for pollution tolerance is included later and tolerance assignments for individuals are included in Attachment B. In general, the number of taxa collected within the annual monitoring stations appears to be increasing since 2001. Tolerant species dominate all annual stations, followed by moderately tolerant species. Several stations have no intolerant species represented during any sample year, while others have a few.


Figure 2-3. Taxonomic Abundances across the CAWS at Annual Monitoring Stations.

The most frequently observed species across all stations included gizzard shad (Dorosoma cepedianum), common carp (Cyprinus carpio), and largemouth bass (Micropterus salmoides), respectively (Figure 2-4). The most frequently observed species at the annual monitoring stations includes gizzard shad, common carp and pumpkinseed (Lepomis gibbosus), respectively (Figure 2-5). The most numerous observed species within the CAWS included gizzard shad ( $\mathrm{n}=6906$ ), emerald shiner (Notropis atherinoides; $\mathrm{n}=2082$ ) and common carp ( $\mathrm{n}=2055$ ), respectively (Figure 2-6). Eleven species are represented by only a single observation for the 2001-2007 period. Finally, gizzard shad, common carp, and largemouth bass have been observed at all stations during the sample period.

The distribution and abundance of gizzard shad in the CAWS is not unusual for large water systems and Simon and Sanders (1999) suggest not including this species in community structure comparisons as a potential source of bias in analysis. Emerald shiner is commonly found in large rivers and appears to thrive in reservoir systems (Becker 1983), so their numbers and distribution within the CAWS is not unexpected. Common carp are found turbid, warm, large river systems of the Midwest (Becker 1983) and their distribution and abundance in the CAWS is also not surprising. Largemouth bass are also abundant in large rivers of the Midwest (Becker 1983), with a presence expected in the CAWS and serve as a popular recreation target species within the system (Personal communication, Bradley 2008). Pumpkinseed also appears to thrive in impounded systems (Becker 1983) so their numbers and distributions are also not unexpected.


Figure 2-4. Species Observations, by Sample Event at all Monitoring Stations for the 2001-2007 Period.


Figure 2-5. Species Observations, by Sample Event at Annual Monitoring Stations for the 2001-2007 Period.


Figure 2-6. Total Number of Individuals Collected during the 2001-2007 Sample Period (black bars are referenced to the axis on the left, blue bars are referenced to the axis on the right).

Figure 2-7 describes the sample collections, among years at the annual monitoring stations. The graphs depict the variation of samples collected at the stations among years. Meador and McIntyre (2003) observed high variability (Coefficient of Variation (CV) = 0.23 ) in species richness from boat electrofished samples of up to $+/-5$ species and concluded that their variation resulted from sampling efficiency, rather than environmental variation. However, they noted that increased variation among years at a station was related to increasing station depth (Meador and McIntyre 2003). Paller (1995) suggests that a CV of 0.20 is the maximum desirable level of variability in catch per unit effort for electrofishing. The CV for the CAWS annual monitoring stations ranged from $0.16-0.4$. The Lockport station had the highest CV (0.40), while the Cal-Sag station at Cicero Avenue had the lowest CV (0.16). The high CV at the Lockport station may be related to the site conditions of confined, deep channels, no access to shallow water areas and a species community that is dominated by mobile species such as gizzard shad, carp, and a range of sunfishes. These findings are also consistent with Meador and McIntyre (2003) in their descriptions of highly variable non-wadeable sites.

Finally, in 2007, the District deployed Fyke nets as a supplemental sampling method for three stations within the CAWS. The Fyke net collected data was compared to the closest electrofishing event in space and time in an attempt to understand how this additional collection method may be of value for use in the CAWS fishery monitoring program for capturing smaller age-class fish. Fyke nets are selective for migratory fish that follow shorelines (Hubert 1996). The 2007 samples resulted in relatively small catches compared to electrofishing and seemed biased towards smaller size classes (Figures 2-8, 2-9, 2-10). The Cal-Sag at Harlem Avenue resulted in the largest catch of 34 individuals. Of the 34 individuals, only four bluegill (total length ranging $31-37 \mathrm{~mm}$ ) were collected with the remaining species being minnows. Only three individuals were collected at the Cal-Sag at Cicero Avenue site: two minnow and one bluegill (total length 31 mm ). The Cal-Sag at Southwest Highway site found 11 individuals: 7 bluegill (total length 23-46 mm), one green sunfish (total length 48 mm ), and the remaining were minnows. Overall, the catch total lengths from the Fyke net samples ranged between 23 mm and 66 mm . Little can be drawn from the small catches of the 2007 Fyke net sample data other than the samples seemed biased towards small samples of young, potentially year 1 (Becker 1983) bluegill and minnows. Future, alternative approaches may include light-traps that target young-of-year fishes to try to understand reproduction within various portions of the CAWS.




Figure 2-7. 2001 to 2007 Annual Station Fish Survey Results




Figure 2-7. 2001 to 2007 Annual Station Fish Survey Results - Continued


Figure 2-6. 2001 to 2007 Annual Station Fish Survey Results - Continued


Figure 2-8. Results of Electrofishing and Fyke Net Samples by Length Interval, from 2007 Samples near Harlem Avenue on the Cal-Sag Channel.


Figure 2-9. Results of Electrofishing and Fyke Net Samples by Length Interval , from 2007 Samples near Southwest Highway on the Cal-Sag Channel.


Figure 2-10. Results of Electrofishing and Fyke Net Samples by Length Interval, from 2007 Samples near Cicero Avenue on the Cal-Sag Channel.

## 3. SELECTION OF FISH METRICS

Fish metric selection and calculation is a common form of fish data analysis (Flotemersch et al. 2006). The general approach for screening fish metrics to determine which will be most useful and appropriate for this study follows methods applied in development of fish IBIs, as documented in peer-reviewed scientific literature. As stated in the preceding section, the objective of this study is not to develop a new IBI for the CAWS, but the process of metric development involves review, analysis, and reduction of fish metrics, so the methods used in the literature to develop IBIs provides a sound basis for screening of metrics appropriate for the CAWS.

### 3.1 COMPILATION OF FISH METRICS

Roset et al. (2007) suggests that starting with a large list of relevant candidate metrics increases the rigor of the system-specific metric selection process, by removing a level of a priori bias retained from previous studies. Lyons et al. (2001) provides a list of 26 fish metrics that were used as the starting point for the Wisconsin large warm water river IBI. The Lyons study is particularly relevant because it was developed in the Midwest for a range of larger river types, it is frequently cited, and Lyons' methodology is well-documented. Starting with Lyons' list of 26 fish metrics, LimnoTech then reviewed other relevant and significant IBI documents to identify other potentially applicable metrics:

- The Illinois IBI (IDNR 2000) was consulted as it currently provides the reference that the Illinois Environmental Protection Agency (IEPA) uses to determine attainment with aquatic life uses and may offer applicable metrics for the unique conditions within the CAWS (IEPA 2005). From this reference, ten additional metrics were added.
- The Ohio Boatable IBI (OEPA 1988) was consulted because it is frequently cited, still used after 20 years, one of the few fish IBI developed specifically for nonwadeable waters in the Midwest, and may offer applicable metrics for the unique conditions within the CAWS. Three additional metrics were included from the Ohio IBI.
- Karr's original work (Karr 1981) on fish IBIs was consulted because it was the seminal work on fish IBIs and most subsequent fish IBI work has been derived from it. No additional metrics were identified from this reference because they are included, as appropriate, in the above IBIs.

The metrics from the Illinois and Ohio IBIs increased the total number of metrics under consideration to 40 . In addition to these previously used metrics, review of fish data from the CAWS and knowledge of the system suggested that some additional metrics would be worthy of consideration, including the following:

- Percent intolerant species by number and by weight - These metrics were added to provide additional quantification of the prevalence of pollution-sensitive individuals. This may provide information beyond the number of intolerant species.
- Percent moderately tolerant species by number and by weight - Previous studies have grouped species into tolerant or intolerant categories, however modifications to water quality standards recently proposed by the Illinois EPA have used the term "intermediately tolerant", so the inclusion of metrics that reflect species that are moderately tolerant to water quality impacts may be useful.
- Number of tolerant species - This metric was included to provide a metric of direct comparison with the number of intolerant and moderately tolerant species.
- Number of sunfish species, excluding largemouth bass - This metric was added because sunfish metrics used in other IBIs either included all sunfish or excluded both smallmouth and largemouth bass. Because smallmouth bass are a cool water species and are less tolerant of anthropogenic impacts, it was desirable to include them, while excluding largemouth bass because of their wide distribution across the CAWS.

With the addition of these 'custom' metrics, the list of potential fish metrics for consideration in this Study totaled 46. Review of additional scientific literature did not identify any more applicable metrics for inclusion, suggesting that the starting metric list will provide the rigor suggested by Roset et al. (2007). The 46 fish metrics and their sources are listed in Table 3.1.

Table 3-1. Initial List of Fish Metrics.

| Fish Metric | Metric Name | Source |
| :---: | :---: | :---: |
| \%DELT_(n) | \% Diseased or with eroded fins, lesions, or tumors | Lyons et al. (2001) |
| CPUE | catch per unit effort |  |
| WPUE | weight per unit effort | " |
| \%LRIV_(n) | \% large river species by count | " |
| \%LRIV_(wt) | \% large river species by weight | " |
| \%RIV_(n) | \% riverine species by count | " |
| \%RIV_(wt) | \% riverine species by weight | " |
| \%RNDSCK_(n) | \% round sucker species by count | " |
| \%RNDSCK_(wt) | \% round sucker species by weight | " |
| \%TOL_( n ) | \% tolerant species by count | " |
| \%TOL_(wt) | \% tolerant species by weight | " |
| INT | number of intolerant species | " |
| RIV | number of riverine species | " |
| \%LTHPL_( n ) | \% lithophilic spawners by count | " |
| \%LTHPL_(wt) | \% lithophilic spawners by weight | " |
| NAT | number of native species | " |
| SCKR | number of sucker species | " |
| SR | total number of species | " |
| SUN1 | number of sunfish species, excluding smallmouth and largemouth bass | " |
| SUN2 | number of sunfish species, including smallmouth and largemouth bass | " |
| \%INSCT_(n) | \% insectivores by count | " |
| \%INSCT_(wt) | \% insectivores by weight | " |
| \%OMV_(n) | \% omnivores by count | " |
| \%OMV_(wt) | \% omnivores by weight | " |
| \%TC_(n) | \% top carnivores by count | " |
| \%TC_(wt) | \% top carnivores by weight | " |
| PRTOL | proportion of Illinois tolerant species | IDNR, 2000 |
| LITOT | IL ratio of non tolerant coarse-mineral-substrate spawners |  |
| INTOL | number of IL native intolerant species | " |
| NFSH | number of IL native species | " |
| NMIN | number of IL native minnow species | " |
| NSUC | number of IL native sucker species | " |
| NSUN | number of IL native sunfish species | " |
| GEN | IL ratio of generalist feeders | " |
| NBINV | IL native benthic invertivore species | " |
| SBI | IL ratio of specialist benthic invertivore species | " |
| TNI | total number of individuals | OEPA, 1988 |
| OH_B_Sun | number of OH native sunfish species | " |
| \%OH_B_OMN(n) | \% OH omnivores, excluding channel catfish | " |
| \%INT_(n) | \% intolerant species by count | New for this Study |
| \%INT_(wt) | \% intolerant species by weight | " |
| \%MOD_(n) | \% moderately intolerant species by count | " |
| \%MOD_(wt) | \% moderately intolerant species by weight | " |
| MOD | number of moderately tolerant species | " |
| TOL | number of tolerant species | " |
| SUN3 | number of sunfish species, excluding largemouth bass | " |

### 3.2 SPECIFICATION OF TOLERANCE VALUES

Several of the metrics identified for screening are intended to be relative indicators of species tolerance to pollution and other human impacts. Therefore these metrics require that species be classified according to their pollution tolerance. This is significant because proposed water quality standards for the CAWS are defined in terms of maintaining aquatic-life populations of fish species that are tolerant, intermediately tolerant, and/or intolerant. It should be noted that the proposed water quality standards do not assign fish species to these tolerance categories, nor do they refer to sources from which to derive tolerance assignments.

The classification of fishes into tolerance categories has typically been based on best professional judgment (BPJ) assignments of species based on general responses to environmental degradation (Meador and Carlisle 2007). Meador and Carlisle (2007) cited that the relative success of BPJ classifications of tolerance in the Midwest may be a result of the perceived homogeneity of regional conditions and that the assignments may have limited geographic application. Further, tolerance assignments rarely discriminate among pollutant stressors. Meador and Carlisle (2007) found that stressors such as suspended sediment, conductivity, chloride and total phosphorus provided a better measure of pollution tolerance assignment than the typically considered stressors of temperature, dissolved oxygen and pH. For example, white sucker (Catostomus commersoni) and fathead minnow (Pimephales promelas) are generally categorized as tolerant to pollutants by Illinois DNR (IDNR 2000) and Meador and Carlisle (2007), despite their intolerance to low dissolved oxygen and high temperatures (Meador and Carlisle 2007).
Unfortunately, the detailed stressor assignments by Meador and Carlisle (2007) have not been developed for the Midwest region and do not consider many of the CAWS species, so their method will not be used here, but warrants future consideration.

The approach for assigning CAWS species to pollution tolerance categories of tolerant, intolerant or moderately tolerant, attempted to rely on locally derived sources, although no single source covered all species found within the CAWS. The approach started with tolerance assignments established at the state level (IDNR 2008), then for the Midwest (Lyons et al. 2001), at the national level (Meador and Carlisle 2008) and then for specific references where a species was not included in the previous documents.

The State of Illinois has developed a manual for calculating fish IBIs that is in draft form with continued updates (IDNR 2000). The manual includes pollution tolerance assignments for a range of species. The IDNR (2000) assignments only include tolerant or intolerant for those with any assignment and most species in the state list have no assignment (that is, they are given a "-"). The classifications were derived from regional fish manuals including Smith (1979), Becker (1983), Karr et al. (1986), Jenkins and Burkhead (1994), Bertrand et al. 1996, OEPA (1988) and BPJ, where information was not available (IDNR 2008). These classifications were retained as a primary reference sources.

The next level of tolerance assignment was derived from Lyons et al. (2001). The Lyons paper provided additional assignments to some species not assigned by IDNR
(2008) but also restricted species assignments to tolerant or intolerant categories only, with the remaining species assigned as "other". The tolerance assignments of Lyons stems from his earlier paper (Lyons 1992) where three qualitative criteria are used:

1) a known high degree of sensitivity to the types of environmental degradation as described by Becker (1983) and other regional fish publications;
2) areas of observed regions of decline in Wisconsin where environmental problems are known; and
3) designations used in other IBIs.

Meador and Carlisle (2007) from the U.S. Geological Survey (USGS) conducted an extensive analysis and assignment of numerous species into tolerant, moderately tolerant, and intolerant categories based on a recently published, quantified evaluation against physiochemical variables. The data set used for this effort is from the USGS national program and collected data from the USGS National Water Quality Assessment Program. These assignments were applied after the Lyons assignments. This effort resulted in a database of tolerance assignments for most remaining fish species, except for some remaining exotics. Finally, for those species not given tolerance assignments by the aforementioned efforts, species-specific papers were consulted and referenced for final pollution tolerance assignments. The tolerance values assigned for each species are included in Attachment B.

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## 4. SCREENING OF FISH METRICS

The procedures and rationale for screening of fish metrics are described below.

### 4.1 SCREENING OBJECTIVES

The process of screening the fish metrics had two primary objectives, as described below:

1. First, it was necessary to reduce the list of fish metrics to a more manageable number. Because the data corresponding to these metrics will be used for comparison to water quality and habitat data, too large a number of fish metrics would be too cumbersome. Metrics used to assess fishes vary based on the physical and biotic nature of the system (Flotemersch et al. 2006). Most fish IBIs reviewed for this study used a final set of ten to sixteen metrics (Karr 1981; OEPA 1988; Hughes et al., 1998; IDNR 2000; Lyons et al., 2001), so the goal was to reduce the list to within this range.
2. Second, the current scientific literature suggests that it is important to retain at least one metric from each major category of ecological function: species richness and composition, indicator species, trophic function, reproductive function, and individual abundance and condition (Simon and Lyons 1995; Lyons et al. 2001; Roset et al. 2007). Each category reflects a different aspect of fish assemblages that responds uniquely to aquatic ecosystem stressors (Hughes and Oberdorff 1999).

With these objectives in mind, the initial list of fish metrics was screened using the process described in the following sections.

### 4.2 METRICS LACKING DATA

The initial step in the screening process was to identify metrics for which there were no data available. This was essential, because the metrics will eventually be used for statistical or other quantitative comparisons to other data types (i.e., water quality and habitat) and the lack of data would preclude such quantitative comparisons.

Review of the CAWS fish data from 2001 to 2007 revealed two metrics for which no data exist in the CAWS: the percentage of round sucker taxa (genera Cycleptus, Hypentelium, Minytrema, and Moxostoma) by weight and by number (\%RNDSCK_(n) and \%RNDSCK_(wt)). Based on this observation, these metrics were eliminated from further consideration. This initial screening reduced the list of fish metrics from 46 to 44 .

### 4.3 METRIC RANGE

Review of the scientific literature for fish IBI development shows that a typical method of screening fish metrics is to examine those metrics that reflect the number of species identified in a particular category or type and to screen out those that represent relatively
few species (McCormick et al. 2001; Emery et al. 2003). This so-called "range test" is used to eliminate metrics for which between 0 and 2 species were identified.

The "range test" was applied to the CAWS fish data and four metrics were found for which only one or two species were identified between 2001 and 2007. These four metrics were: the number of Illinois native benthic invertivore species (NBINV), the number of Illinois native sucker species (NSUC), the number of sucker species (SCKR), and the Illinois ratio of specialist benthic invertivore species (SBI). On the basis of this observation, these four metrics were eliminated from further consideration, reducing the number of potential metrics to 40 .

### 4.4 METRIC REDUNDANCY

A very common method of screening metrics is to analyze the metrics for redundancy with each other. This method of screening is commonly used in index development (Hughes et al. 1998; Lyons et al. 2001, Emery et al. 2003, Wilhelm et al. 2005). In this analysis, Pearson's correlation was calculated for pairs of metrics and the resulting correlation values were used to screen out statistically redundant metrics. This process is described in more detail below.

Before calculating the Pearson correlation coefficients, the metrics were evaluated for normality and several metrics were found to have skewed distributions. Some were right skewed, others were left skewed. The left skewed metrics were log transformed, resulting in near-normal distributions and include the following metrics: WPUE, TNI, TOL_TNI, CPUE, \%TC_(wt), \%LTHPL_(wt), \%MOD_(n), and \%MOD_(wt). For the right skewed metrics (mostly data representing proportions) the arcsine-square-root-transform was evaluated, but because the distribution shapes did not improve these metrics were left untransformed.

Pearson correlation coefficients were calculated between the individual metrics in order to identify metrics that are highly correlated. Correlated metrics indicate some degree of redundancy, i.e. they respond similarly to characteristics of the CAWS system and can be used to derive similar conclusions. Threshold correlation strength had to be chosen to identify the metrics with "strong" correlation, as reported in the literature. In the literature reviewed, this threshold correlation value was usually between 0.6 and 0.75 (Lyons et al. 2001; McCormick et al. 2001; Emery et al. 2003; Whittier et al. 2007). For this analysis a value of 0.6 was used, which is what Lyons used for his large warm water river IBI (Lyons et al. 2001). Thus, pairs of metrics with a correlation coefficient above the threshold were defined as redundant and only one metric of the pair was retained for subsequent analyses. The matrix of Pearson's correlation coefficients is presented in Attachment C.

Because many metrics were highly correlated with multiple metrics, some judgment was necessary in using this screening method to insure representation from each of the five ecological function categories. For example, the original list of 46 metrics only contained
three reproductive function metrics and five abundance and condition metrics, therefore these metrics were, in some cases, preferentially retained.

This screening step was successful in reducing the number of metrics from 40 to 16 . The list of metrics remaining after screening for redundancy is presented in Table 4-1.

Table 4-1. Fish Metrics Remaining after Screening for Redundancy.

| Fish Metric | Metric Name | Ecological <br> Function Category |
| :--- | :--- | :---: |
| \%DELT_(n) | \% Diseased or with eroded fins, lesions, or tumors | ACM |
| CPUE | catch per unit effort | ACM |
| \%LTHPL_(n) | \% lithophilic spawners by count | RFM |
| \%LTHPL_(wt) | \% lithophilic spawners by weight | RFM |
| \%INSCT_(n) | \% insectivores by count | TFM |
| \%INSCT_(wt) | \% insectivores by weight | TFM |
| \%TC_(n) | \% top carnivores by count | TFM |
| \%TC_(wt) | \% top carnivores by weight | TFM |
| PRTOL | proportion of Illinois tolerant species | ISM |
| LITOT | IL ratio of non tolerant coarse-mineral-substrate spawners | RFM |
| INTOL | number of IL native intolerant species | ISM |
| NMIN | number of IL native minnow species | SRC |
| NSUN | number of IL native sunfish species | SRC |
| GEN | IL ratio of generalist feeders | TFM |
| \%INT_(n) | \% intolerant species by count | ISM |
| \%MOD_(wt) | \% moderately intolerant species by weight | ISM |

### 4.5 METRIC VARIABILITY

After applying the methods described above, the number of retained metrics (16) still exceeded the target number of metrics, so the retained metrics were inspected to determine whether a rational scientific basis could be identified for elimination of any of them.

It was noted that the set of metrics listed in Table 4-1 contained three pairs of metrics that represented similar fish attributes for both count and weight:

- \% lithophilic spawners by count (\%LTHPL_(n)) and weight (\%LTHPL_(wt))
- \% insectivores by count (\%INSCT_(n)) and weight (\%INSCT_(wt))
- \% top carnivores count (\%TC_(n)) and weight (\%TC_(wt))

In addition, two metrics remained that represented intolerant species: \%INT_(n) and INTOL. Because each of these four pairs of metrics measure the same attributes of fish assemblages, it seemed appropriate to select one metric from each pair to carry forward. To determine which metric in each pair to retain, the variability of the metrics within the data set was examined. The rationale for using metric variability as a screening measure was that preference should be given to metrics that exhibited greater variation within the
system, since those metrics will be more likely to help identify relationships to other system attributes such as water quality and physical habitat.

Calculated values for each of the paired metrics were extracted from the CAWS fish database and the coefficient of variation (CV) for each metric was calculated using all data from each year from 2001 through 2007 to give a measure of data variability in each year for each metric. The CV for each metric was also calculated at each of seven annual sampling stations for all years to determine variability across the system. The results are discussed below.

The system-wide CVs for \%LTHPL_(wt) and \%LTHPL_(n) are depicted graphically in Figure 4-1.


Figure 4-1. Coefficient of Variation for \%LTHPL_(wt) and \%LTHPL_(n), for 2001 through 2007 Data.

Although the CVs for both \%LTHPL_(wt) and \%LTHPL_(n) are both very low (less than 0.5 in every year), the calculated value for $\%$ LTHPL_( $^{-}$) is consistently higher, in many cases double that of \%LTHPL_(wt). The CVs for \%LTHPL_( n ) also appear to exhibit more variability over time than for \%LTHPL_(wt), which is also evident from the CVs calculated for the annual sampling stations depicted in Figure 4-2. Based on these observations, \%LTHPL_(n) was retained and \%LTHPL_(wt) was eliminated.


Figure 4-2. Coefficient of Variation for \%LTHPL_(wt) and \%LTHPL_(n) at Annual Sampling Stations.

The same comparison was made for \%INSCT_(n) and \%INSCT_(wt). In this case, the CV for \%INSCT_(n) is consistently higher than for \%INSCT_(wt), both on a systemwide basis across multiple years (Figure 4-3) as well as when compared between annual sampling stations (Figure 4-4). On the basis of these comparisons, \%INSCT_(n) was retained and \%INSCT_(wt) was eliminated.


Figure 4-3. Coefficient of Variation for \%INSCT_(wt) and \% INSCT_(n), for 2001 through 2007 Data.


Figure 4-4. Coefficient of Variation for \%INSCT_(wt) and \% INSCT_(n) at Annual Sampling Stations.

Similarly, the CVs for \%TC_(n) and \%TC_(wt) were compared. In the case of this metric pair, most of the CVs for \%TC_(wt) were above 1.0, while all the CVs for \%TC_(n) were below 1.0, suggesting that \%TC_(wt) has significantly higher variability (Figure 4-5).
While some sampling stations exhibited similar CVs for both \%TC_(wt) and \%TC_(n) (Figure 4-6), three stations had significantly higher CVs for \%TC_(wt). Based on these observations, \%TC_(wt) was retained and \%TC_(n) was eliminated.

Finally, the CVs for \%INT_(n) and INTOL were compared both on a system-wide basis for each sampling year and for each annual sampling station across all years. The comparison of system-wide variability through time (Figure 4-7) clearly indicates that \%INT_(n) has higher variability than INTOL, even though the inter-station comparison (Figure 4-8) shows similarity between the two metrics in terms of variability. On the basis of these observations, \%INT_(n) was retained and INTOL was eliminated.


Figure 4-5. Coefficient of Variation for \%TC_(wt) and \% TC_(n), for 2001 through 2007 Data.


Figure 4-6. Coefficient of Variation for \%TC_(wt) and \% TC_(n) at Annual Sampling Stations.


Figure 4-7. Coefficient of Variation for \%INT_(n) and INTOL, for 2001 through 2007 Data.


Figure 4-8. Coefficient of Variation for \%INT_(n) and INTOL at Annual Sampling Stations.

In summary, based on review of metric variability as quantified by each metric's coefficient of variation, the following metric selections were made:

- \%LTHPL_(n) was retained over \%LTHPL_(wt);
- \%INSCT_(n) was retained over \%INSCT_(wt);
- \%TC_(wt) was retained over \%TC_(n); and
- $\%$ INT_(n) was retained over INTOL.

These selections reduced the list of metrics to 12 , which are summarized in the following section.

## 5. FINAL RECOMMENDED LIST OF METRICS

After completion of the screening process described in the preceding section, twelve metrics were retained for use in the CAWS (Table 5-1). The retained metrics are representative of each of the five ecological function categories as recommended by Simon and Lyons (1995), Lyons et al. (2001), Roset et al. (2007): species richness and composition (SRC), indicator species (ISM), trophic function (TFM), reproductive function (RFM), and individual abundance and condition (ACM). These are further described below,

SRC category includes two native species metrics. Species richness and composition are a measure of species diversity and Hughes and Oberdorff (1999) suggest using native species metrics for assessing physical or water quality stressors where non-natives are abundant, as found in the CAWS. Both metrics are also used by the State of Illinois and should be appropriate measures for species richness assessments within the CAWS.

ISM includes three proportional metrics of tolerant, moderately tolerant and intolerant measures. Proportional measures for species have been recommended by others as well (Karr et al. 1986; Lyons et al. 1995). The current numbers of intolerant species across the CAWS is generally low and it is generally expected that the proportion of intolerant species is responding to physical and water quality stressors unique to the CAWS. However, it is expected that these species would respond positively to stressor reductions and may provide an appropriate metric for the CAWS. Both tolerant and moderately tolerant species are wide-spread across the CAWS and it is assumed that the tolerant metrics would respond negatively to physical and water quality improvements while moderately tolerant species proportions increase with the reduction of stressors. All three proportional measures are applicable measures across the CAWS.

TFM includes a range of feeding metrics for the CAWS that include top carnivores, generalists and insect feeders. It is generally expected that top carnivores and insectivores would respond negatively to physical and water quality stressors, while generalists would respond positively to these stressors (Flotemersch et al. 2006). All three metrics are applicable across the CAWS, are appropriate measures of trophic function and are supported by the original work of Karr (1981) and subsequent authors (Hughes and Oberdorff 1999).

RFM includes a proportion of all lithophilic species as well as intolerant lithophilic species native to Illinois. It is generally expected that lithophilic species would respond negatively to both physical and water quality stressors (Flotemersch et al. 2006). Although it is expected that lithophilic habitat is limited across the CAWS, these metrics are included because existing habitat conditions as well as future improvements within portions of the CAWS should result in a positive response by these metrics. The metrics are used within the Illinois IBI as
well as others (Emery et al. 1999; Flotemersch et al. 2006) and are appropriate for the CAWS.

ACM includes a metric for the condition of the sampled fishes as well as the efficiency of the collection methods. It is generally expected that the observed number of physical anomalies of collected fishes changes in response to a range of water quality stressors. Hughes and Oberdorff (1999) suggest including this metric where the possibility for changes in the incidence of disease and deformity exist. Hughes and Oberdorff (1999) describe sample abundance as a surrogate for system productivity but caution that nutrient and thermal enrichment may affect this metric response. Typically, it is expected that the efficiency of collected fishes decrease in response to both water quality and habitat stressors (Flotemersch et al. 2006) but the uniqueness of the CAWS conditions may warrant special consideration of the use of this metric in subsequent analysis. Both measures are commonly used measures for ACM and are appropriate for the CAWS.

In summary, the methods used for fish metric selection for the CAWS are appropriate, literature supported and robust methods. These methods have produced a final metric list that is appropriate and sensitive to responses of both physical habitat and water quality conditions within the CAWS and will be useful for further fish-habitat and fish-water quality analyses.

Table 5-1. Final Recommended Fish Metrics for Use in the CAWS.

| Fish Metric | Metric Name | Ecological <br> Function Category |
| :--- | :--- | :---: |
| \%DELT_(n) | \% Diseased or with eroded fins, lesions, or tumors | ACM |
| CPUE | catch per unit effort | ACM |
| \%LTHPL_( $n$ ) | \% lithophilic spawners by count | RFM |
| \%INSCT_( $n$ ) | \% insectivores by count | TFM |
| \%TC_(wt) | \% top carnivores by weight | TFM |
| PRTOL | proportion of Illinois tolerant species | ISM |
| LITOT | IL ratio of non tolerant coarse-mineral-substrate spawners | RFM |
| NMIN | number of IL native minnow species | SRC |
| NSUN | number of IL native sunfish species | SRC |
| GEN | IL ratio of generalist feeders | TFM |
| \%INT_(n) | \% intolerant species by count | ISM |
| \%MOD_(wt) | \% moderately intolerant species by weight | ISM |

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## ATTACHMENT A:

## CAWS FISH DATA STATIONS AND SAMPLING DATES <br> USED IN THIS ANALYSIS

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| IEPA Station Description | Station ID | Station Description | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | Total Count |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| North Shore Channel | 35 | North Shore Channel at Central Street | 9/24/01 |  |  |  | 7/20/05 |  |  | 2 |
|  | 36 | North Shore Channel at Touhy Avenue | 9/26/01 | 7/31/02 | 7/24/03 | 9/29/04 | 7/21/05 | 7/10/06 | 7/12/07 | 7 |
|  | 101 | North Shore Channel at Foster Avenue | 9/27/01 |  |  |  | 9/8/05 |  |  | 2 |
|  | 102 | North Shore Channel at Oakton Street | 9/25/01 |  |  |  | 7/20/05 |  |  | 2 |
| North Branch Chicago River from its confluence with North Shore Channel to the south end of the North Avenue Turning Basin | 37 | North Branch Chicago River at Wilson Avenue | 10/1/01 |  |  |  | 9/7105 |  |  | 2 |
|  | 73 | North Branch Chicago River at Diversey Parkway | 10/3/01 |  |  |  | 9/6/05 |  |  | 2 |
| Little Calumet River from its confluence with Calumet River and Grand Calumet River to its confluence with Calumet-Sag Channel | 56 | Little Calumet River at Indiana Avenue |  |  | 9/29/03 |  |  |  | 7/30/07 | 2 |
|  | 76 | Little Calumet River at Halsted Street | 9/12/01 | 9/16/02 | 9/29/03 | 9/30/04 | 9/27/05 | 7/21/06 | 7/31/07 | 7 |
|  | 902,SEPA2 | Little Calumet River at SEPA 2 |  |  |  |  | 10/20/05 | 10/20/06 |  | 2 |
| Calumet-Sag Channel | 43 | Calumet-Sag Channel at Route 83 |  |  | 7/30/03 |  |  |  | 9/14/07 | 2 |
|  | 58 | Calumet-Sag Channel at Ashland Avenue |  |  | 9/5/03 |  |  |  | 8/1/07 | 2 |
|  | 59 | Calumet-Sag Channel at Cicero Avenue | 9/14/01 | 9/17/02 | 7/31/03 | 8/31/04 | 9/29/05 | 7/24/06 | 8/2107 | 7 |
|  | 903,SEPA3 | Calumet-Sag Channel at SEPA 3 |  |  | 10/6/03 |  | 10/20/05 |  | 10/31/07 | 3 |
|  | 904,SEPA4 | Calumet-Sag Channel at SEPA 4 |  |  | 10/3/03 | 10/19/04 | 10/18/05 | 10/30/06 | 10/29/07 | 5 |
|  | 905,SEPA5 |  |  |  | 10/1/03 | 10/18/04 | 10/18/05 | 10/17/06 | 10/23/07 | 5 |
|  |  | Calumet-Sag Channel at 104th Street |  |  |  |  |  |  | 9/14/07 | 1 |
|  |  | Calumet-Sag Channel at Kedzie Avenue |  |  |  |  |  |  | 9/13/07 | 1 |
|  |  | Calumet-Sag Channel at Southwest Highway |  |  |  |  |  |  | 9/13/07 | 1 |
| North Branch Chicago River from the south end of the North Avenue Turning Basin to its confluence with South Branch Chicago River and Chicago River | 46 | North Branch Chicago River at Grand Avenue | 10/2/01 | 8/1/02 | 7/23/03 | 8/27/04 | 7/18/05 | 7/11/06 | 7/11/07 | 7 |
| Chicago River | 74 | Chicago River at Lake Shore Drive |  | 8/2/02 |  |  |  | 7/26/06 |  | 2 |
|  | 100 | Chicago River at Wells Street |  | 8/21/02 |  |  |  | 7/27/06 |  | 2 |
| South Branch Chicago River and its South Fork | 39 | South Branch Chicago River at Madison Street |  | 8/27/02 |  |  |  | 7/28/06 |  | 2 |
|  | 40 | Chicago Sanitary and Ship Canal at Damen Avenue |  | 8/19/02 |  |  |  | 8/30/06 |  | 2 |
|  | 99 | Bubbly Creek at Archer Avenue |  | 8/20/02 |  |  |  | 9/5/06 |  | 2 |
|  | 108 | South Branch Chicago River at Loomis Street |  | 8/26/02 |  |  |  | 9/12/06 |  | 2 |
|  | 99.2 | Bubbly Creek at 35th St. |  |  | 9/30/03 | 10/20/04 | 8/10/05 |  |  | 3 |
|  | 99.1 |  |  |  | 9/30/03 | 10/20/04 | 8/10/05 |  |  | 3 |
|  | 99.3 | Bubbly Creek at RAPS |  |  | 9/30/03 | 10/20/04 | 8/10/05 |  |  | 3 |
| Chicago Sanitary and Ship Canal | 41 | Chicago Sanitary and Ship Canal at Harlem Avenue | 9/7101 | 9/3/02 | 7/21/03 | 8/24/04 | 8/26/05 | 8/21/06 | 7/16/07 | 7 |
|  | 42 | Chicago Sanitary and Ship Canal at Route 83 |  | 8/28/02 |  |  |  | 8/31/06 |  | 2 |
|  | 48 | Chicago Sanitary and Ship Canal at Stephen Street |  | 9/10/02 |  |  |  | 8/28/06 |  | 2 |
|  | 75 | Chicago Sanitary and Ship Canal at Cicero Avenue** | 9/4/01 | 8/29/02 | 7/18/03 | 8/23/04 | 8/22/05 | 8/29/06 | 7/17/07 | 7 |
|  | 92 | Chicago Sanitary and Ship Canal at Lockport (16th Street) | 9/4/01 | 9/11/02 | 7/29/03 | 8/30/04 | 9/15/05 | 7/25/06 | 7/10/07 | 7 |
|  | 905.1,SEPA5_CSSC | Chicago Sanitary and Ship Canal at SEPA 5 |  |  | 10/1/03 | 10/18/04 | 10/18/05 | 10/17/06 | 10/23/07 | 5 |
|  | Total Stations: 34 |  | 12 | 15 | 17 | 13 | 20 | 19 | 17 | 113 |

## ATTACHMENT B:

## LIST OF FISH SPECIES IDENITIFIED IN THE CAWS (2001-2007) AND THEIR TOLERANCE ASSIGNMENTS

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| Scientific Name | Common Name | Tolerant | Intolerant | Moderate |
| :---: | :---: | :---: | :---: | :---: |
| Alosa pseudoharengus | alewife | $\mathrm{X}^{4}$ |  |  |
| Alosa chrysochloris | skipjack herring |  |  | $x^{5}$ |
| Dorosoma cepedianum | gizzard shad | $\mathrm{x}^{3}$ |  |  |
| Oncorhynchus mykiss | rainbow trout |  | $\mathrm{x}^{3}$ |  |
| Umbra limi | central mudminnow | $\mathrm{x}^{2}$ |  |  |
| Esox lucius | northern pike |  |  | $\mathrm{x}^{3}$ |
| Carassius auratus | goldfish | $\mathrm{X}^{1}$ |  |  |
| Cyprinus carpio | common carp | $\mathrm{X}^{1}$ |  |  |
| Notemigonus crysoleucas | golden shiner | $\mathrm{X}^{1}$ |  |  |
| Semotilus atromaculatus | creek chub | $\mathrm{X}^{1}$ |  |  |
| Cyprinella spiloptera | spotfin shiner |  |  | $\mathrm{x}^{3}$ |
| Pimephales promelas | fathead minnow | $\mathrm{x}^{1}$ |  |  |
| Pimephales notatus | bluntnose minnow | $\mathrm{x}^{1}$ |  |  |
| Notropis atherinoides | emerald shiner | $\mathrm{x}^{3}$ |  |  |
| Notropis hudsonius | spottail shiner |  | $\mathrm{X}^{2}$ |  |
| Notropis stramineus | sand shiner |  |  | $\mathrm{X}^{3}$ |
| Ictiobus niger | black buffalo |  | $\mathrm{X}^{2}$ |  |
| Catostomus commersoni | white sucker | $\mathrm{x}^{1}$ |  |  |
| Ictalurus punctatus | channel catfish | $\mathrm{X}^{3}$ |  |  |
| Ameiurus natalis | yellow bullhead | $\mathrm{x}^{1}$ |  |  |
| Ameiurus melas | black bullhead |  |  | $\mathrm{x}^{3}$ |
| Ameiurus nebulosus | brown bullhead |  |  | $\mathrm{x}^{3}$ |
| Noturus gyrinus | tadpole madtom |  |  | $\mathrm{x}^{3}$ |
| Fundulus notatus | blackstripe topminnow | $\mathrm{x}^{3}$ |  |  |
| Gambusia affinis | mosquitofish | $x^{3}$ |  |  |
| Labidesthes sicculus | brook silverside |  |  | $\mathrm{x}^{3}$ |
| Morone saxatilis | striped bass |  |  | $\mathrm{x}^{6}$ |
| Morone chrysops | white bass | $\mathrm{x}^{3}$ |  |  |
| Morone mississippiensis | yellow bass |  |  | $\mathrm{X}^{9}$ |
| Morone americana | white perch | $\mathrm{x}^{7}$ |  |  |
| Pomoxis nigromaculatus | black crappie | $\mathrm{x}^{3}$ |  |  |
| Pomoxis annularis | white crappie | $\mathrm{x}^{3}$ |  |  |
| Ambloplites rupestris | rock bass |  | $\mathrm{x}^{2}$ |  |
| Micropterus salmoides | largemouth bass | $\mathrm{x}^{3}$ |  |  |
| Micropterus dolomieu | smallmouth bass |  | $\mathrm{X}^{1}$ |  |
| Lepomis gulosus | warmouth |  |  | $\mathrm{x}^{3}$ |
| Lepomis cyanellus | green sunfish | $\mathrm{X}^{1}$ |  |  |
| Lepomis macrochirus | bluegill |  |  | $\mathrm{X}^{3}$ |
| Lepomis gibbosus | pumpkinseed |  |  | $\mathrm{x}^{3}$ |
| Lepomis humilis | orangespotted sunfish | $\mathrm{x}^{3}$ |  |  |
| Stizostedion vitreum | walleye |  |  | $\mathrm{x}^{3}$ |


| Scientific Name | Common Name | Tolerant | Intolerant | Moderate |
| :--- | :--- | :--- | :--- | :---: |
| Perca flavescens | yellow perch |  |  | $\mathrm{X}^{3}$ |
| Aplodinotus grunniens | freshwater drum | $\mathrm{X}^{3}$ |  |  |
| Neogobius melanostomus | round goby | $\mathrm{X}^{8}$ |  |  |
| Cyprinus spp. | carp x goldfish | $\mathrm{X}^{11}$ |  |  |
| Oncorhynchus tshawytscha | chinook salmon |  | $\mathrm{X}^{9}$ |  |
| Oncorhynchus kisutch | coho salmon |  | $\mathrm{X}^{9}$ |  |
| Lepomis spp. | green sunfish $\times$ bluegill | $\mathrm{X}^{11}$ |  |  |
| Lepomis spp. | green sunfish $\times$ longear | $\mathrm{X}^{11}$ |  |  |
| Lepomis spp. | green sunfish $\times$ pumpkinseed | $\mathrm{X}^{11}$ |  |  |
| Oreochromis niloticus | nile tilapia | $\mathrm{X}^{10}$ |  |  |
| Lepomis spp. | pumpkinseed x bluegill | $\mathrm{X}^{11}$ |  |  |

## References

$X^{1}$ - IDNR 2000
$X^{2}$ - Lyons et al. 2001
$X^{3}$ - USGS 2008
$X^{4}$ - FWS 1986
$X^{5}$ - Barbour et al. 1999
$X^{6}$ - EPA 2008
$X^{7}$ - FWS 1983
$X^{8}$ - Corkum et al. 2004
$X^{9}$ - Plafkin et al. 1989
$\mathrm{X}^{10}$ - Popma and Masser 1999
$X^{11}$ - LTI 2008

## ATTACHMENT C:

## MATRIX OF PEARSON'S CORRELATION VALUES FOR FISH METRICS

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## Pearson Coefficient Matrix

|  | \%DELT_(n) | \%INSCT_( n ) | \%INSCT_(wt) | \%INT_(n) | \%INT_(wt) | \%LRIV_(n) | \%LRIV_(wt) | \%LTHPL_( n ) | \%LTHPL_(wt | \%MOD_( n ) | \%MOD_(wt) | PH_B_OMV | \%OMV_( n ) | \%OMV_(wt) | \%RIV_(n) | \%RIV_(wt) | \%TC_(n) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \%DELT_( n ) | 1 | 0.265 | 0.351 | 0.036 | -0.177 | -0.053 | 0.287 | -0.010 | 0.321 | 0.128 | -0.203 | 0.306 | 0.313 | 0.323 | 0.012 | 0.280 | 0.237 |
| \%INSCT_(n) | 0.265 | 1 | 0.557 | 0.162 | 0.144 | -0.527 | -0.133 | -0.522 | -0.028 | 0.336 | 0.230 | -0.385 | -0.384 | 0.005 | -0.416 | -0.049 | 0.504 |
| \%INSCT_(wt) | 0.351 | 0.557 | 1 | -0.062 | 0.000 | 0.021 | 0.425 | 0.022 | 0.409 | 0.087 | -0.141 | 0.072 | 0.078 | 0.483 | 0.055 | 0.434 | 0.351 |
| \%INT_(n) | 0.036 | 0.162 | -0.062 | 1 | 0.269 | -0.195 | -0.249 | -0.111 | -0.168 | 0.108 | 0.111 | -0.105 | -0.111 | -0.171 | -0.147 | -0.186 | 0.194 |
| \%INT_(wt) | -0.177 | 0.144 | 0.000 | 0.269 | 1 | 0.078 | -0.321 | 0.121 | -0.098 | -0.077 | 0.056 | -0.362 | -0.363 | -0.347 | 0.057 | -0.316 | -0.002 |
| \%LRIV_(n) | -0.053 | -0.527 | 0.021 | -0.195 | 0.078 | 1 | 0.582 | 0.955 | 0.515 | -0.426 | -0.494 | 0.371 | 0.377 | 0.407 | 0.873 | 0.482 | -0.517 |
| \%LRIV_(wt) | 0.287 | -0.133 | 0.425 | -0.249 | -0.321 | 0.582 | 1 | 0.563 | 0.888 | -0.167 | -0.513 | 0.518 | 0.523 | 0.906 | 0.574 | 0.900 | -0.185 |
| \%LTHPL_(n) | -0.010 | -0.522 | 0.022 | -0.111 | 0.121 | 0.955 | 0.563 | 1 | 0.530 | -0.471 | -0.543 | 0.328 | 0.329 | 0.395 | 0.826 | 0.458 | -0.489 |
| \%LTHPL_(wt) | 0.321 | -0.028 | 0.409 | -0.168 | -0.098 | 0.515 | 0.888 | 0.530 | 1 | -0.128 | -0.485 | 0.421 | 0.410 | 0.808 | 0.521 | 0.810 | -0.166 |
| \%MOD_( n ) | 0.128 | 0.336 | 0.087 | 0.108 | -0.077 | -0.426 | -0.167 | -0.471 | -0.128 | 1 | 0.589 | 0.041 | 0.026 | -0.049 | -0.107 | -0.024 | 0.581 |
| \%MOD_(wt) | -0.203 | 0.230 | -0.141 | 0.111 | 0.056 | -0.494 | -0.513 | -0.543 | -0.485 | 0.589 | 1 | -0.236 | -0.237 | -0.407 | -0.249 | -0.322 | 0.386 |
| \%OH_B_OMV_(n) | 0.306 | -0.385 | 0.072 | -0.105 | -0.362 | 0.371 | 0.518 | 0.328 | 0.421 | 0.041 | -0.236 | 1 | 0.997 | 0.513 | 0.375 | 0.476 | -0.099 |
| \%OMV_( n ) | 0.313 | -0.384 | 0.078 | -0.111 | -0.363 | 0.377 | 0.523 | 0.329 | 0.410 | 0.026 | -0.237 | 0.997 | 1 | 0.518 | 0.376 | 0.479 | -0.090 |
| \%OMV_(wt) | 0.323 | 0.005 | 0.483 | -0.171 | -0.347 | 0.407 | 0.906 | 0.395 | 0.808 | -0.049 | -0.407 | 0.513 | 0.518 | 1 | 0.425 | 0.831 | -0.066 |
| \%RIV_(n) | 0.012 | -0.416 | 0.055 | -0.147 | 0.057 | 0.873 | 0.574 | 0.826 | 0.521 | -0.107 | -0.249 | 0.375 | 0.376 | 0.425 | 1 | 0.609 | -0.277 |
| \%RIV_(wt) | 0.280 | -0.049 | 0.434 | -0.186 | -0.316 | 0.482 | 0.900 | 0.458 | 0.810 | -0.024 | -0.322 | 0.476 | 0.479 | 0.831 | 0.609 | 1 | -0.070 |
| \%TC_(n) | 0.237 | 0.504 | 0.351 | 0.194 | -0.002 | -0.517 | -0.185 | -0.489 | -0.166 | 0.581 | 0.386 | -0.099 | -0.090 | -0.066 | -0.277 | -0.070 | 1 |
| \%TC_(wt) | -0.352 | 0.096 | -0.101 | 0.167 | 0.245 | -0.274 | -0.536 | -0.283 | -0.581 | 0.159 | 0.539 | -0.344 | -0.333 | -0.523 | -0.215 | -0.447 | 0.455 |
| \%TOL_( n ) | 0.015 | -0.230 | 0.247 | -0.132 | 0.057 | 0.625 | 0.499 | 0.708 | 0.455 | -0.672 | -0.591 | 0.320 | 0.330 | 0.434 | 0.385 | 0.372 | -0.296 |
| \%TOL_(wt) | 0.203 | 0.028 | 0.466 | -0.061 | -0.255 | 0.428 | 0.726 | 0.443 | 0.636 | -0.119 | -0.374 | 0.508 | 0.512 | 0.740 | 0.371 | 0.551 | 0.040 |
| CPUE | -0.072 | 0.210 | 0.068 | 0.164 | 0.102 | -0.224 | -0.064 | -0.236 | -0.013 | 0.794 | 0.485 | -0.045 | -0.059 | 0.035 | 0.035 | 0.025 | 0.409 |
| GEN | 0.313 | -0.384 | 0.078 | -0.110 | -0.363 | 0.377 | 0.523 | 0.329 | 0.411 | 0.026 | -0.237 | 0.997 | 1.000 | 0.518 | 0.376 | 0.480 | -0.090 |
| INT | -0.083 | 0.112 | 0.017 | 0.596 | 0.420 | 0.096 | -0.099 | 0.174 | 0.001 | -0.079 | -0.013 | -0.255 | -0.259 | -0.094 | 0.092 | -0.082 | 0.084 |
| INTOL | 0.016 | 0.125 | 0.021 | 0.647 | 0.223 | 0.031 | -0.043 | 0.107 | 0.032 | -0.029 | 0.020 | -0.161 | -0.168 | -0.024 | 0.044 | -0.018 | 0.130 |
| LITOT | 0.049 | 0.131 | -0.066 | 0.983 | 0.154 | -0.189 | -0.198 | -0.113 | -0.146 | 0.111 | 0.094 | -0.066 | -0.072 | -0.124 | -0.143 | -0.141 | 0.192 |
| MOD | -0.114 | 0.129 | 0.071 | 0.041 | 0.117 | -0.035 | -0.003 | -0.029 | 0.023 | 0.570 | 0.345 | -0.032 | -0.047 | 0.073 | 0.126 | 0.069 | 0.251 |
| NAT | -0.122 | 0.184 | 0.183 | 0.113 | 0.155 | -0.028 | 0.056 | 0.033 | 0.077 | 0.312 | 0.180 | -0.073 | -0.082 | 0.132 | 0.068 | 0.114 | 0.301 |
| NFSH | -0.122 | 0.184 | 0.183 | 0.113 | 0.155 | -0.028 | 0.056 | 0.033 | 0.077 | 0.312 | 0.180 | -0.073 | -0.082 | 0.132 | 0.068 | 0.114 | 0.301 |
| NMIN | -0.275 | 0.156 | 0.146 | 0.014 | 0.240 | 0.101 | 0.088 | 0.135 | 0.115 | 0.077 | 0.115 | -0.174 | -0.186 | 0.071 | 0.147 | 0.141 | -0.022 |
| NSUN | 0.065 | 0.152 | 0.072 | 0.347 | 0.000 | -0.169 | -0.087 | -0.116 | -0.068 | 0.473 | 0.206 | 0.078 | 0.064 | 0.063 | -0.056 | -0.063 | 0.474 |
| OH_B_SUN | 0.091 | 0.123 | 0.029 | 0.285 | -0.038 | -0.215 | -0.096 | -0.166 | -0.081 | 0.544 | 0.240 | 0.151 | 0.135 | 0.063 | -0.073 | -0.045 | 0.404 |
| PRTOL | 0.191 | 0.089 | 0.382 | -0.294 | -0.141 | 0.154 | 0.354 | 0.181 | 0.297 | -0.401 | -0.266 | 0.267 | 0.285 | 0.331 | 0.065 | 0.317 | 0.006 |
| RIV | -0.057 | 0.099 | 0.224 | 0.018 | 0.100 | 0.179 | 0.276 | 0.230 | 0.269 | 0.233 | 0.000 | -0.039 | -0.046 | 0.284 | 0.270 | 0.300 | 0.204 |
| SR | -0.107 | 0.183 | 0.220 | 0.077 | 0.186 | 0.025 | 0.097 | 0.096 | 0.128 | 0.282 | 0.100 | -0.104 | -0.115 | 0.155 | 0.096 | 0.127 | 0.274 |
| SUN1 | 0.091 | 0.123 | 0.029 | 0.285 | -0.038 | -0.215 | -0.096 | -0.166 | -0.081 | 0.544 | 0.240 | 0.151 | 0.135 | 0.063 | -0.073 | -0.045 | 0.404 |
| SUN2 | 0.065 | 0.152 | 0.072 | 0.347 | 0.000 | -0.169 | -0.087 | -0.116 | -0.068 | 0.473 | 0.206 | 0.078 | 0.064 | 0.063 | -0.056 | -0.063 | 0.474 |
| SUN3 | 0.088 | 0.134 | 0.057 | 0.350 | -0.026 | -0.175 | -0.063 | -0.118 | -0.047 | 0.487 | 0.203 | 0.100 | 0.083 | 0.083 | -0.047 | -0.019 | 0.398 |
| TNI | -0.187 | -0.100 | 0.129 | 0.016 | 0.187 | 0.356 | 0.337 | 0.390 | 0.345 | 0.093 | -0.027 | 0.107 | 0.096 | 0.377 | 0.434 | 0.341 | -0.072 |
| TOL | -0.076 | 0.167 | 0.264 | -0.044 | 0.110 | 0.029 | 0.156 | 0.107 | 0.165 | 0.142 | -0.019 | -0.073 | -0.079 | 0.200 | 0.054 | 0.160 | 0.241 |
| WPUE | -0.052 | 0.238 | 0.057 | 0.058 | 0.394 | -0.172 | -0.151 | -0.135 | -0.032 | 0.390 | 0.411 | -0.271 | -0.276 | -0.101 | 0.004 | -0.098 | 0.273 |

NOTE: This matrix does not include: 1. \%RNDSCK_(n) or \%RNDSCK_(wt) because there were no data for these metrics (no fish identified in these categories).
2. NBINV, NSUC, and SCKR because each of these species metrics had insufficient numbers of species associated with them in the dataset (2 or fewer)

|  | \%TC_(wt) | \%TOL_(n) | \%TOL_(wt) | CPUE | GEN | INT | INTOL | LITOT | MOD | NAT | NFSH | NMIN | NSUN | OH_B_SUN | PRTOL | RIV | SR | SUN1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \%DELT_( n ) | -0.352 | 0.015 | 0.203 | -0.072 | 0.313 | -0.083 | 0.016 | 0.049 | -0.114 | -0.122 | -0.122 | -0.275 | 0.065 | 0.091 | 0.191 | -0.057 | -0.107 | 0.091 |
| \%INSCT_(n) | 0.096 | -0.230 | 0.028 | 0.210 | -0.384 | 0.112 | 0.125 | 0.131 | 0.129 | 0.184 | 0.184 | 0.156 | 0.152 | 0.123 | 0.089 | 0.099 | 0.183 | 0.123 |
| \%INSCT_(wt) | -0.101 | 0.247 | 0.466 | 0.068 | 0.078 | 0.017 | 0.021 | -0.066 | 0.071 | 0.183 | 0.183 | 0.146 | 0.072 | 0.029 | 0.382 | 0.224 | 0.220 | 0.029 |
| \%INT_(n) | 0.167 | -0.132 | -0.061 | 0.164 | -0.110 | 0.596 | 0.647 | 0.983 | 0.041 | 0.113 | 0.113 | 0.014 | 0.347 | 0.285 | -0.294 | 0.018 | 0.077 | 0.285 |
| \%INT_(wt) | 0.245 | 0.057 | -0.255 | 0.102 | -0.363 | 0.420 | 0.223 | 0.154 | 0.117 | 0.155 | 0.155 | 0.240 | 0.000 | -0.038 | -0.141 | 0.100 | 0.186 | -0.038 |
| \%LRIV_(n) | -0.274 | 0.625 | 0.428 | -0.224 | 0.377 | 0.096 | 0.031 | -0.189 | -0.035 | -0.028 | -0.028 | 0.101 | -0.169 | -0.215 | 0.154 | 0.179 | 0.025 | -0.215 |
| \%LRIV_(wt) | -0.536 | 0.499 | 0.726 | -0.064 | 0.523 | -0.099 | -0.043 | -0.198 | -0.003 | 0.056 | 0.056 | 0.088 | -0.087 | -0.096 | 0.354 | 0.276 | 0.097 | -0.096 |
| \%LTHPL_( $n$ ) | -0.283 | 0.708 | 0.443 | -0.236 | 0.329 | 0.174 | 0.107 | -0.113 | -0.029 | 0.033 | 0.033 | 0.135 | -0.116 | -0.166 | 0.181 | 0.230 | 0.096 | -0.166 |
| \%LTHPL_(wt) | -0.581 | 0.455 | 0.636 | -0.013 | 0.411 | 0.001 | 0.032 | -0.146 | 0.023 | 0.077 | 0.077 | 0.115 | -0.068 | -0.081 | 0.297 | 0.269 | 0.128 | -0.081 |
| \%MOD_(n) | 0.159 | -0.672 | -0.119 | 0.794 | 0.026 | -0.079 | -0.029 | 0.111 | 0.570 | 0.312 | 0.312 | 0.077 | 0.473 | 0.544 | -0.401 | 0.233 | 0.282 | 0.544 |
| \%MOD_(wt) | 0.539 | -0.591 | -0.374 | 0.485 | -0.237 | -0.013 | 0.020 | 0.094 | 0.345 | 0.180 | 0.180 | 0.115 | 0.206 | 0.240 | -0.266 | 0.000 | 0.100 | 0.240 |
| \%OH_B_OMV_( n ) | -0.344 | 0.320 | 0.508 | -0.045 | 0.997 | -0.255 | -0.161 | -0.066 | -0.032 | -0.073 | -0.073 | -0.174 | 0.078 | 0.151 | 0.267 | -0.039 | -0.104 | 0.151 |
| \%OMV_(n) | -0.333 | 0.330 | 0.512 | -0.059 | 1.000 | -0.259 | -0.168 | -0.072 | -0.047 | -0.082 | -0.082 | -0.186 | 0.064 | 0.135 | 0.285 | -0.046 | -0.115 | 0.135 |
| \%OMV_(wt) | -0.523 | 0.434 | 0.740 | 0.035 | 0.518 | -0.094 | -0.024 | -0.124 | 0.073 | 0.132 | 0.132 | 0.071 | 0.063 | 0.063 | 0.331 | 0.284 | 0.155 | 0.063 |
| \%RIV_(n) | -0.215 | 0.385 | 0.371 | 0.035 | 0.376 | 0.092 | 0.044 | -0.143 | 0.126 | 0.068 | 0.068 | 0.147 | -0.056 | -0.073 | 0.065 | 0.270 | 0.096 | -0.073 |
| \%RIV_(wt) | -0.447 | 0.372 | 0.551 | 0.025 | 0.480 | -0.082 | -0.018 | -0.141 | 0.069 | 0.114 | 0.114 | 0.141 | -0.063 | -0.045 | 0.317 | 0.300 | 0.127 | -0.045 |
| \%TC_( n ) | 0.455 | -0.296 | 0.040 | 0.409 | -0.090 | 0.084 | 0.130 | 0.192 | 0.251 | 0.301 | 0.301 | -0.022 | 0.474 | 0.404 | 0.006 | 0.204 | 0.274 | 0.404 |
| \%TC_(wt) | 1 | -0.128 | -0.153 | 0.179 | -0.333 | 0.194 | 0.145 | 0.134 | 0.167 | 0.260 | 0.260 | 0.068 | 0.302 | 0.224 | 0.002 | 0.121 | 0.216 | 0.224 |
| \%TOL_( n ) | -0.128 | 1 | 0.641 | -0.504 | 0.330 | 0.061 | 0.016 | -0.131 | -0.274 | -0.027 | -0.027 | 0.056 | -0.182 | -0.236 | 0.625 | 0.099 | 0.024 | -0.236 |
| \%TOL_(wt) | -0.153 | 0.641 | 1 | -0.046 | 0.512 | -0.044 | 0.000 | -0.026 | -0.012 | 0.102 | 0.102 | 0.012 | 0.141 | 0.096 | 0.563 | 0.214 | 0.142 | 0.096 |
| CPUE | 0.179 | -0.504 | -0.046 | 1 | -0.060 | 0.205 | 0.203 | 0.137 | 0.781 | 0.658 | 0.658 | 0.435 | 0.584 | 0.634 | -0.448 | 0.567 | 0.647 | 0.634 |
| GEN | -0.333 | 0.330 | 0.512 | -0.060 | 1 | -0.259 | -0.168 | -0.071 | -0.047 | -0.082 | -0.082 | -0.186 | 0.064 | 0.135 | 0.286 | -0.047 | -0.116 | 0.135 |
| INT | 0.194 | 0.061 | -0.044 | 0.205 | -0.259 | 1 | 0.932 | 0.529 | 0.156 | 0.342 | 0.342 | 0.218 | 0.376 | 0.221 | -0.354 | 0.366 | 0.369 | 0.221 |
| INTOL | 0.145 | 0.016 | 0.000 | 0.203 | -0.168 | 0.932 | 1 | 0.602 | 0.142 | 0.348 | 0.348 | 0.196 | 0.448 | 0.297 | -0.332 | 0.361 | 0.338 | 0.297 |
| LITOT | 0.134 | -0.131 | -0.026 | 0.137 | -0.071 | 0.529 | 0.602 | 1 | 0.001 | 0.057 | 0.057 | -0.056 | 0.343 | 0.280 | -0.280 | -0.009 | 0.022 | 0.280 |
| MOD | 0.167 | -0.274 | -0.012 | 0.781 | -0.047 | 0.156 | 0.142 | 0.001 | 1 | 0.775 | 0.775 | 0.567 | 0.544 | 0.571 | -0.495 | 0.629 | 0.752 | 0.571 |
| NAT | 0.260 | -0.027 | 0.102 | 0.658 | -0.082 | 0.342 | 0.348 | 0.057 | 0.775 | 1 | 1.000 | 0.747 | 0.670 | 0.642 | -0.172 | 0.869 | 0.967 | 0.642 |
| NFSH | 0.260 | -0.027 | 0.102 | 0.658 | -0.082 | 0.342 | 0.348 | 0.057 | 0.775 | 1.000 | 1 | 0.747 | 0.670 | 0.642 | -0.172 | 0.869 | 0.967 | 0.642 |
| NMIN | 0.068 | 0.056 | 0.012 | 0.435 | -0.186 | 0.218 | 0.196 | -0.056 | 0.567 | 0.747 | 0.747 | 1 | 0.170 | 0.167 | -0.085 | 0.687 | 0.720 | 0.167 |
| NSUN | 0.302 | -0.182 | 0.141 | 0.584 | 0.064 | 0.376 | 0.448 | 0.343 | 0.544 | 0.670 | 0.670 | 0.170 | 1 | 0.940 | -0.293 | 0.563 | 0.645 | 0.940 |
| OH_B_SUN | 0.224 | -0.236 | 0.096 | 0.634 | 0.135 | 0.221 | 0.297 | 0.280 | 0.571 | 0.642 | 0.642 | 0.167 | 0.940 | 1 | -0.284 | 0.527 | 0.606 | 1.000 |
| PRTOL | 0.002 | 0.625 | 0.563 | -0.448 | 0.286 | -0.354 | -0.332 | -0.280 | -0.495 | -0.172 | -0.172 | -0.085 | -0.293 | -0.284 | 1 | -0.099 | -0.139 | -0.284 |
| RIV | 0.121 | 0.099 | 0.214 | 0.567 | -0.047 | 0.366 | 0.361 | -0.009 | 0.629 | 0.869 | 0.869 | 0.687 | 0.563 | 0.527 | -0.099 | 1 | 0.901 | 0.527 |
| SR | 0.216 | 0.024 | 0.142 | 0.647 | -0.116 | 0.369 | 0.338 | 0.022 | 0.752 | 0.967 | 0.967 | 0.720 | 0.645 | 0.606 | -0.139 | 0.901 | 1 | 0.606 |
| SUN1 | 0.224 | -0.236 | 0.096 | 0.634 | 0.135 | 0.221 | 0.297 | 0.280 | 0.571 | 0.642 | 0.642 | 0.167 | 0.940 | 1.000 | -0.284 | 0.527 | 0.606 | 1 |
| SUN2 | 0.302 | -0.182 | 0.141 | 0.584 | 0.064 | 0.376 | 0.448 | 0.343 | 0.544 | 0.670 | 0.670 | 0.170 | 1.000 | 0.940 | -0.293 | 0.563 | 0.645 | 0.940 |
| SUN3 | 0.222 | -0.206 | 0.110 | 0.609 | 0.083 | 0.368 | 0.449 | 0.349 | 0.549 | 0.658 | 0.658 | 0.174 | 0.970 | 0.970 | -0.323 | 0.573 | 0.632 | 0.970 |
| TNI | 0.020 | 0.230 | 0.337 | 0.567 | 0.095 | 0.285 | 0.234 | -0.019 | 0.578 | 0.730 | 0.730 | 0.612 | 0.427 | 0.434 | -0.041 | 0.733 | 0.765 | 0.434 |
| TOL | 0.177 | 0.146 | 0.210 | 0.484 | -0.080 | 0.216 | 0.195 | -0.086 | 0.538 | 0.895 | 0.895 | 0.679 | 0.553 | 0.521 | 0.113 | 0.865 | 0.945 | 0.521 |
| WPUE | 0.240 | -0.296 | -0.154 | 0.693 | -0.277 | 0.356 | 0.251 | -0.018 | 0.612 | 0.588 | 0.588 | 0.461 | 0.318 | 0.325 | -0.274 | 0.488 | 0.609 | 0.325 |

## Pearson Coefficient Matrix

|  | SUN2 | SUN3 | TNI | TOL | WPUE |
| :--- | :---: | :---: | :---: | :---: | :---: |
| \%DELT_(n) | 0.065 | 0.088 | -0.187 | -0.076 | -0.052 |
| \%INSCT_(n) | 0.152 | 0.134 | -0.100 | 0.167 | 0.238 |
| \%INSCT_(wt) | 0.072 | 0.057 | 0.129 | 0.264 | 0.057 |
| \%INT_(n) | 0.347 | 0.350 | 0.016 | -0.044 | 0.058 |
| \%INT_(wt) | 0.000 | -0.026 | 0.187 | 0.110 | 0.394 |
| \%LRIV_(n) | -0.169 | -0.175 | 0.356 | 0.029 | -0.172 |
| \%LRIV_(wt) | -0.087 | -0.063 | 0.337 | 0.156 | -0.151 |
| \%LTHPL_(n) | -0.116 | -0.118 | 0.390 | 0.107 | -0.135 |
| \%LTHPL_(wt) | -0.068 | -0.047 | 0.345 | 0.165 | -0.032 |
| \%MOD_(n) | 0.473 | 0.487 | 0.093 | 0.142 | 0.390 |
| \%MOD_(wt) | 0.206 | 0.203 | -0.027 | -0.019 | 0.411 |
| \%OH_B_OMV_(n) | 0.078 | 0.100 | 0.107 | -0.073 | -0.271 |
| \%OMV_(n) | 0.064 | 0.083 | 0.096 | -0.079 | -0.276 |
| \%OMV_(wt) | 0.063 | 0.083 | 0.377 | 0.200 | -0.101 |
| \%RIV_(n) | -0.056 | -0.047 | 0.434 | 0.054 | 0.004 |
| \%RIV_(wt) | -0.063 | -0.019 | 0.341 | 0.160 | -0.098 |
| \%TC_(n) | 0.474 | 0.398 | -0.072 | 0.241 | 0.273 |
| \%TC_(wt) | 0.302 | 0.222 | 0.020 | 0.177 | 0.240 |
| \%TOL_(n) | -0.182 | -0.206 | 0.230 | 0.146 | -0.296 |
| \%TOL_(wt) | 0.141 | 0.110 | 0.337 | 0.210 | -0.154 |
| CPUE | 0.584 | $\mathbf{0 . 6 0 9}$ | 0.567 | 0.484 | $\mathbf{0 . 6 9 3}$ |
| GEN | 0.064 | 0.083 | 0.095 | -0.080 | -0.277 |
| INT | 0.376 | 0.368 | 0.285 | 0.216 | 0.356 |
| INTOL | 0.448 | $\mathbf{0 . 4 4 9}$ | 0.234 | 0.195 | 0.251 |
| LITOT | 0.343 | 0.349 | -0.019 | -0.086 | -0.018 |
| MOD | 0.544 | 0.549 | 0.578 | 0.538 | $\mathbf{0 . 6 1 2}$ |
| NAT | $\mathbf{0 . 6 7 0}$ | $\mathbf{0 . 6 5 8}$ | $\mathbf{0 . 7 3 0}$ | $\mathbf{0 . 8 9 5}$ | 0.588 |
| NFSH | $\mathbf{0 . 6 7 0}$ | $\mathbf{0 . 6 5 8}$ | $\mathbf{0 . 7 3 0}$ | $\mathbf{0 . 8 9 5}$ | 0.588 |
| NMIN | 0.170 | $\mathbf{0 . 1 7 4}$ | $\mathbf{0 . 6 1 2}$ | $\mathbf{0 . 6 7 9}$ | 0.461 |
| NSUN | $\mathbf{1 . 0 0 0}$ | $\mathbf{0 . 9 7 0}$ | 0.427 | 0.553 | 0.318 |
| OH_B_SUN | $\mathbf{0 . 9 4 0}$ | $\mathbf{0 . 9 7 0}$ | 0.434 | 0.521 | 0.325 |
| PRTOL | -0.293 | -0.323 | -0.041 | 0.113 | -0.274 |
| RIV | 0.563 | $\mathbf{0 . 5 7 3}$ | $\mathbf{0 . 7 3 3}$ | $\mathbf{0 . 8 6 5}$ | 0.488 |
| SR | $\mathbf{0 . 6 4 5}$ | $\mathbf{0 . 6 3 2}$ | $\mathbf{0 . 7 6 5}$ | $\mathbf{0 . 9 4 5}$ | $\mathbf{0 . 6 0 9}$ |
| SUN1 | $\mathbf{0 . 9 4 0}$ | $\mathbf{0 . 9 7 0}$ | 0.434 | 0.521 | 0.325 |
| SUN2 | $\mathbf{1}$ | $\mathbf{0 . 9 7 0}$ | 0.427 | 0.553 | 0.318 |
| SUN3 | $\mathbf{0 . 9 7 0}$ | $\mathbf{1}$ | 0.445 | 0.534 | 0.327 |
| TNI | 0.427 | 0.445 | $\mathbf{1}$ | $\mathbf{0 . 7 2 1}$ | 0.511 |
| TOL | 0.553 | $\mathbf{0 . 5 3 4}$ | $\mathbf{0 . 7 2 1}$ | $\mathbf{1}$ | 0.477 |
| WPUE | 0.318 | 0.327 | 0.511 | 0.477 | $\mathbf{1}$ |
|  |  |  |  |  |  |


[^0]:    ${ }^{1}$ In 2007, the District supplemented fish collections with Fyke net samples but, because this method is not consistent with other methods, these data were not included in this analysis.

